Negative Correlation between Urinary Selenium and Arsenic Levels of the Residents Living in an Arsenic-Contaminated Area in Bangladesh

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Toxicological and metabolic interactions of selenium (Se) with arsenic (As) have been reported in many experimental studies. However, for human populations, possible interactions between As and Se and their toxicological significance have not been established. In this study, we have examined the relationship between Se and As in spot urine samples collected from the inhabitants of two rural communities of northeast Bangladesh. The urinary As (U_{Ae}) and Se (U_{Se}) concentrations were determined by hydride generation atomic absorption spectrometry (HGAAS). Negative correlation between U_{As} and U_{Se} was found in both males and females. Considering the exposure condition in these communities, it is unlikely that the correlation reflected the correlation between the intakes of these elements. This result suggests that As exposure from drinking water or food may change Se metabolism in humans. Possible toxicological significance of the interaction is discussed.

Key words — arsenic, selenium, interaction, Bangladesh

INTRODUCTION

The effects on human health of arsenic (As) exposure from drinking tubewell water of geologic origin are a major public health problem in many countries, including Bangladesh, India, China, Chile, Argentina, and Peru. This hazard is particularly se-

rious in Bangladesh, where 35 to 77 million of the 125 million inhabitants are estimated to be at risk.¹) Chronic exposure to As causes dermatological, cardiovascular, neurological, and neurodevelopmental disorders, and cancers.^{2,3})

Selenium (Se) is an essential microelement that negates the toxic effects of several heavy metals,⁴⁾ and its toxicological and metabolic interactions with As have been reported in many animal experiments. Moxon and Du Bois⁵⁾ revealed the protective effects of As against liver damage and growth depression, both of which are caused by excessive consumption of seleniferous grain, though the mechanisms of protection by Se against As toxicity have not yet been clarified. The experiments of rats injected with ⁷⁵Se and As have demonstrated that increase in the As dose markedly increased the amount of ⁷⁵Se in the excreted bile, and the increased amount of 75Se was approximately equivalent to the decreased amount of ⁷⁵Se in the liver.^{6,7)} Keyon *et al.*⁸⁾ have reported that mice fed Se-deficient diet eliminated arsenate, arsenite and dimethylarsinic acid (DMA) in urine more slowly than those fed Se-suficient diet.

For human populations, very few such studies have been carried out. Lin and Yang⁹⁾ reported that black foot disease (BFD) patients in Taiwan had lower blood and hair Se concentrations than normal controls. Wang¹⁰⁾ indicated that urinary Se (U_{Se}) concentrations were significantly lower in BFD patients than in healthy controls. Wuyi *et al.*¹¹⁾ found a decrease in the hair, urinary, and blood As concentrations, together with marked improvement in the skin manifestations, in patients supplemented with Se (100–200 μ g Se/day) versus those receiving a placebo in a 14-month study of arsenism patients. The toxicological significance of Se and its possible in-

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teractions with As have not been established in human pupulations. It is possible that As could disturb normal metabolism of Se in human populations. This study examined the relationship between As and Se levels in urine in two Bangladeshi communities with known As contamination.¹²

MATERIALS AND METHODS

Study Populations and Sample Collection – The communities, (SV and SP), located in Sibganj thana, Nawabganj district, northwestern Bangladesh were the subject areas in this study. Their water supply is entirely dependent on about 100 tubewells in the two villages, some of which were installed by individual households while others were installed by the government for common use among several households. Spot urine samples were collected from 170 males and 259 females aged 12 years or more in February and March 1999, along with the collection of all tubewell water.¹²⁾ The participation of the inhabitants was voluntary. The collected urine samples were kept in a freezer and transported to Japan, where they were kept at -80°C until element determination. All the procedures were approved by the Ethics Committee of the Graduate School of Medicine of the University of Tokyo, and written informed consent was obtained from each participant.

Urinary As and Se Determination — The total As concentration in the urine samples was determined using an atomic absorption spectrometer equipped with a flow-injection hydride generator [hydride generation atomic absorption spectrometry (HGAAS); ZL-4100, Perkin Elmer, U.S.A.] after wet-ashing the samples with a mixture of HNO₃, HClO₄, and H₂SO₄.¹²⁾ To determine Se concentration, the urine samples were wet-ashed by heating with a mixture of HNO₃ and HClO₄, and Se was determined with the HGAAS after prereduction with HCl. The detection limits of As and Se were 3 and 1 $\mu g/l$, respectively.

The accuracy of the As and Se assays was ensured by including reference material (NIES#18 "human urine," National Institute of Environmental Studies, Tsukuba, Japan). The values obtained for both elements fell within the certified ranges (0.137 ± 0.011 for As and 0.059 ± 0.005 mg/l for Se). The creatinine (Cre) concentration was spectrophotometrically determined using a commercial kit (Creatinine Wako, Wako Pure Pharmaceuticals,

 Table 1. Urinary Arsenic and Selenium Concentration by Sex and by Village

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village and sex	U _{As} [µg/g Cre]	U _{Se} [µg/g Cre]
SP		
male $(n = 79)$	131.8 ± 2.3	23.4 ± 1.5
	$(32-1822)^{a}$	(4–71)
female ($n = 137$)	166.0 ± 2.5	23.3 ± 1.6
	(24–2017)	(4–63)
SV		
male $(n = 90)$	208.9 ± 1.9	17.4 ± 1.6
	(62–1257)	(4-48)
female (<i>n</i> = 119)	218.8 ± 2.4	16.2 ± 1.8
	(42–1653)	(2-44)
$ANOVA^{b)}$		
village	p < 0.001	p < 0.001
sex	0.05	NS
village x sex	NS	NS
Note: Geometric mean and SD are shown a) The ranges of		

Note: Geometric mean and SD are shown. *a*) The ranges of U_{As} and U_{Se} are indicated in parentheses. *b*) Two-way ANOVA using log-transformed variables. NS, not significant (p > 0.1).

Osaka, Japan), based on Jaffe's reaction.¹³⁾ **Statistical Analysis** — Each individual's U_{Se} and urinary As (U_{As}) concentrations were logarithmtransformed to ensure normality in distribution. Between-group differences in the transformed data were tested by analysis of variance (ANOVA), using the JMP software program (version 4.0; SAS Institute, Cary, NC, U.S.A.). The reported correlation coefficient r is the Pearson correlation coefficient. The probability level was set at p < 0.05.

RESULTS

U_{As} and U_{Se} Concentrations

The U_{As} and U_{Se} concentrations varied considerably among the subjects. The mean was significantly higher in SV than those in SP for both sexes, while the mean U_{Se} was significantly higher in SP than in SV for both sexes (Table 1). With respect to sex differences, U_{As} was higher in females than in males when the two village groups were analyzed separately or together (Table 1).

Relation between U_{As} and U_{Se}

When the data from the two communities were pooled, U_{As} was negatively correlated with U_{Se} in both females (r = -0.25, p < 0.01) and males (r = -0.16, p < 0.05), although a significant correlation was found only in SV females (r = -0.29, p < 0.01; data not shown) when the two communi-

• SP Male Female SV 100 100 USe [µg/g Cre] USe [µg/g Cre] 10 0.16, p<0.0 0.29, p<0.0 10 100 1000 10 100 1000 UAs [µg/g Cre]

Fig. 1. Relationship between U_{As} and U_{Se} in Males (Left) and in Females (Right)

UAs [µg/g Cre]

Each dot denotes one individual; open circles = SP, closed triangles = SV. The lines are the regression lines for each sex with the Pearson's correlation coefficient between $logU_{As}$ and $logU_{Se}$.

ties were analyzed separately (Fig. 1).

DISCUSSION

The results of this study highlight the inverse correlation between U_{As} and U_{Se} in both females and males in the combined populations of the two villages. It should be note that we have observed with a high-performance liquid chromatography-inductively coupled plasma mass spectrometry (HPLC-ICP-MS) assay that the urine samples of this population contained only inorganic arsenic and its methylated metabolites [monomethylarsonic acid (MMA) and DMA] but no other chemical species such as arsenobetaine were detected. Hsuch et al.¹⁴⁾ recently showed that U_{As} increased with U_{Se} among healthy residents of As-affected area (Lanyang Basin) in Taiwan, differing from the observation in this study. The reason for the discrepancy among these observations has not been elucidated, but it might be due to the difference in the levels of As exposure, since the U_{As} concentrations of our subjects (Table 1; note that the arithmetic mean of the U_{As} in our subjects were 281 μ g/g Cre) were higher than those in Hsueh's study (*arithmetic* mean, $96.9 \pm 7.3 \mu g/l$).

The considerable overlap of the ranges of U_{As} and U_{s_e} of the two villages suggests that the inverse correlation was not merely resulting from the between-village differences in these variables. There are several ways to explain this inverse correlation, which will be discussed hereafter.

First, since the urinary excretions of both Se and

As are known to reflect the intakes of each element, the inverse correlation might reflect the inverse correlation between the intakes of these two elements. It is known that the major source of As in this area is tube-well water, whose As concentrations vary from less than 1 to more than 500 μ g/l. In fact, in the present subjects, a highly positive correlation was obtained between As concentration in the tube-well water (TW_{As}) used by an individual and his/her U_{As}. Thus, if the inverse correlation reflected the correlation between the intakes of As and Se, it must be assumed those who had high TW_{As} has lower intake of Se, which should be mainly derived from food. Since TW_{As} is randomly distributed among the subjects, existence of such relationship is unlikely.

Alternatively, the inverse correlation might be indicative of metabolic interaction between these two elements. Although human evidence is scarce, several animal studies showed that such metabolic interactions do exist. In rats, dietary exposure to As significantly enhanced the fecal excretion of Se, with a concomitant decrease in urinary excretion of Se.¹⁵⁾ In rats increasing As dose markedly increased the biliary excretion of co-administered ⁷⁵Se, and the increase was approximately equivalent to the decrease of ⁷⁵Se in the liver.^{6,7)} Using mice, we have also observed that repeated administration of subtoxic dose of As decreased hepatic Se concentration (unpublished data). Although the precise mechanisms are not clear, these observations suggests that As exposure changes Se metabolism, enhancing the excretion of Se via bile or feces, thereby leading to decreased urinary Se excretion. Overall, these experimental observations are in line with the present result in human. In this connection, it is plausible that the between-village difference in As exposure (due to the between-village difference in TW_{As}) led to the between-village difference in U_{se} in the study populations.

In conclusion, this study showed negative correlation between U_{As} and U_{Se} in both males and females of two As-affected rural communities in northwestern Bangladesh, suggesting that As exposure from drinking water or food may change the metabolism of Se in human population. Such disturbance of normal Se metabolism per se may lead to several adverse effects; e.g., impairment of antioxidant selenoenzymes (glutathione peroxidase and thioredoxin reductase). In addition, considering that thioredoxin reductase is a sensitive target of As,¹⁶⁾ the interaction may also potentiate the toxicity of As. In this respect, the relationships between nutritional status regarding Se and the markers of Asinduced oxidative stress or diseases (cancer, skin manifestation etc.) needed to be examined in future epidemiological studies.

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